


# Assessment of marine litter in the mangrove forest in the Ciénaga de Mallorquín, Colombian Caribbean region

## *Evaluación de la basura marina en el bosque de manglar en la Ciénaga de Mallorquín, región Caribe Colombiano*

Anubis Vélez-Mendoza<sup>1</sup>, Carlos Villamil<sup>2</sup>, Karina Castellanos<sup>1</sup> & Yamileth Domínguez-Haydar<sup>1</sup>  
[ajvelez@mail.uniatlantico.edu.co](mailto:ajvelez@mail.uniatlantico.edu.co), [cvillamile@gmail.com](mailto:cvillamile@gmail.com), [karinacastellanos@mail.uniatlantico.edu.co](mailto:karinacastellanos@mail.uniatlantico.edu.co), [yamilethdominguez@mail.uniatlantico.edu.co](mailto:yamilethdominguez@mail.uniatlantico.edu.co)

<sup>1</sup>Universidad del Atlántico, Barranquilla, Colombia , <sup>2</sup>Consultor, Barranquilla, Colombia

Received manuscript: November 10<sup>th</sup> of 2022.  
Received in revised format: May 28<sup>th</sup> of 2023.  
Accepted: June 1<sup>st</sup> of 2023.

### Abstract

The Ciénaga de Mallorquín is a unique coastal lagoon close to Barranquilla City. It is the first study on this topic in the Ciénaga de Mallorquín that evaluated the state of pollution based on marine litter inside the mangrove forest in the Ciénaga de Mallorquín. A total of 860 items and 77.9 kg of litter were collected, equivalent to an average density of 23.89 items\*m<sup>-2</sup> and 2.16 kg\*m<sup>-2</sup>. Plastic was the main litter item (43.55 %), followed by polystyrene, rubber, and processed wood, most of which comes from inland sources such as dumping and recreational and urban activities and is transported mainly by local streams. Of the litter collected, 87.3% was characterized as persistent buoyancy items. The mangrove forest is classified in a dirty state (Class IV), with a considerable number of hazardous items (Class III). The CM needs an urgent, integral, and proactive management plan to reduce litter at its sources, adopting and improving measures such as education and public awareness, good management practices, recycling, and reuse.

*Keywords:* coastal lagoon; plastic pollution; hazardous litter items; Ramsar site.

### Resumen

La Ciénaga de Mallorquín es una laguna costera única cerca de la ciudad de Barranquilla. Este es el primer estudio sobre este tema en la Ciénaga de Mallorquín que evalúa el estado de contaminación basado en basura marina dentro del bosque de manglar en la Ciénaga de Mallorquín. Se recolectaron 860 ítems y 77,9 kg de basura, equivalente a una densidad promedio de 23,89 ítems/m<sup>2</sup> y 2,16 kg/m<sup>2</sup>. El plástico fue el principal rubro (43,55 %), seguido del poliestireno, el caucho y la madera procesada, la mayoría de la cual proviene de fuentes terrestres como vertidos, actividades recreativas y urbanas, y es transportada principalmente por arroyos locales. De la basura recolectada, 87,3% se caracterizó como elementos de flotabilidad persistente. El bosque de manglar está clasificado como Clase IV (estado sucio), con una cantidad considerable de elementos peligrosos (Clase III). El CM necesita un plan de gestión urgente, integral y proactivo para reducir la basura en sus fuentes, adoptando y mejorando medidas como la educación y la sensibilización pública, las buenas prácticas de gestión, el reciclaje y la reutilización.

*Palabras clave:* laguna costera; contaminación plástica; basura peligrosa; sitio Ramsar.

#### **How to cite this article:**

Vélez-Mendoza, A., Villamil, C., Castellanos, K. and Domínguez-Haydar, Y., (2023). Assessment of marine litter in the mangrove forest in the Ciénaga de Mallorquín, Colombian Caribbean region. BOLETÍN DE CIENCIAS DE LA TIERRA, 53, pp. 23 - 37. DOI:<https://doi.org/10.15446/rbct.n53.105749>.

## 1. Introduction

Mangrove forests cover around 13,200,000 ha along tropical and subtropical coasts, with 3,799.54 km<sup>2</sup> distributed along the Pacific Ocean and the Caribbean Sea coasts of Colombia. These are a highly productive ecosystem with a wide range of species, an essential source of food, tourism, medicinal plants, and forestry products (Hamilton and Casey 2016, Villate *et al.* 2020). The ecosystem is made up of plants with aerial roots that enhance gas exchange, nutrient recycling, and water purification. Stabilize plants on unstable substrates protect against erosion and extreme events like cyclones and tsunamis, and act as litter traps (Chong 2005, Díaz 2011, Kauffman *et al.* 2011, Ivar do Sul *et al.* 2014, Norris *et al.* 2017, Martín *et al.* 2019).

The mangrove is a fragile ecosystem, affected by disorganized infrastructure planning, industrial development, uncontrolled tourism, deforestation, fishing activities, sedimentation, climate change, and poor litter management (Hartley *et al.* 2015, Rangel-Buitrago *et al.* 2017). Deforestation resulted in the loss of 35.0 % of the world's mangroves in the span between 1980 and 2000 (Valiela *et al.* 2001, Giesen *et al.* 2007, Kauffman *et al.* 2011) with a range of annual deforestation from 2000 to 2012 of 0.16-0.39 % (Fries *et al.* 2019), and constant litter input is turning the ecosystem into one of the final disposal destinations, contributing to its degradation and converting it into a litter dump (Cordeiro and Costa 2010, Garcés-Ordóñez and Bayona 2019, Riascos *et al.* 2019).

Marine litter (ML) is defined as any solid and persistent materials manufactured or processed that are disposed of or abandoned in the marine and coastal environment (UNEP 2009). They are indirectly transported by seas, rivers, sewage water, stormwater, winds, and fishing activities, or thrown by people into the sea, beaches, and other marine and coastal ecosystems, like the mangroves (Derraik 2002, Gall and Thompson 2015, Rangel-Buitrago *et al.* 2019a, Williams and Rangel-Buitrago *et al.* 2019).

ML (Marine litter) pollution suffocates seedlings by hindering light penetration in the water column, introduces invasive species, induces morbidity of marine species, affects human health (inflammatory lesions, neurodegenerative diseases, immunological disorders, and cancers), and increases negative economic impacts (Boix-Morán 2012, Bulow and Ferdinand 2013, Green *et al.* 2015, Lozoya *et al.* 2016, Antão-Barboza *et al.* 2018, Botterell *et al.* 2019, Garcés-Ordóñez and Bayona 2019).

Colombia produces about 11.6 million tons of litter annually, of which only 17% is recycled (DANE 2018). A few studies have evaluated the presence and impacts of ML

on coastal ecosystems and mangrove forests. In Colombia, ~65.0 % of its coastal populations dispose of their litter in the open air by burning it, burying it in the ground, or dumping it in bodies of water like rivers and lagoons (Garcés-Ordóñez *et al.* 2019). These studies include the mangroves of the Buenaventura Bay on the Pacific coast (Riascos *et al.* 2019) and of the Ciénaga Grande de Santa Marta on the Caribbean coast (Garcés-Ordóñez *et al.* 2019, Garcés-Ordóñez and Bayona 2019).

Preliminary studies carried out by Williams *et al.* (2016a), Gracia *et al.* (2018), and Rangel-Buitrago *et al.* (2017, 2018, 2019a, 2020) refer to the presence, composition, magnitude, fate, and serious environmental problems related to ML along the Colombian Caribbean coast where the Ciénaga de Mallorquín (CM) is located. The CM is part of a Ramsar site declared in 1998 by Decree 0224, updated in 2009 by Decree 3888. This Ramsar site has a high diversity of fish, mammals, amphibians, reptiles, mollusks, crustaceans, and birds (migratory and endemic).

Recently, relevant restoration projects have been carried out in the CM by the current administration of the District of Barranquilla and by an environmental government agency called EPA Barranquilla Verde (Blanchar *et al.* 2020). The ecosystem presents environmental degradation caused by increased sedimentation processes, urbanization of nearby areas, illegal deforestation, industrial litter, domestic sewage water, hospital litter, and litter in general (Vera and De La Rosa Muñoz 2003, GTA 2005, Berrocal *et al.* 2018, Fuentes *et al.* 2018, Chacón *et al.* 2020, Portz *et al.* 2020).

Our study focused on assessing the magnitude, composition, sources, and environmental impacts of ML in the mangrove ecosystem of the CM. As a critical input not only for the recovery and conservation processes of the ecosystem but also to demonstrate the pressing need for formal actions leading to Integrated Coastal Marine Management. It is the first study on this topic in the CM.

## 2. Materials and methods

### 2.1 Study site

The CM is a shallow estuarine coastal lagoon with an area of close to 650 ha and an average depth of 0.90 m, surrounded by floodplains and dunes (Rebolledo-Colina and León-Luna, 2017; Fuentes *et al.*, 2018; Villate *et al.*, 2020). It is located on the Caribbean coast before the mouth of the Magdalena River (75°52'00" W and 11°05'00" N). This ecosystem is in the jurisdiction of the municipalities of Barranquilla City and Puerto Colombia town (Fig. 1).

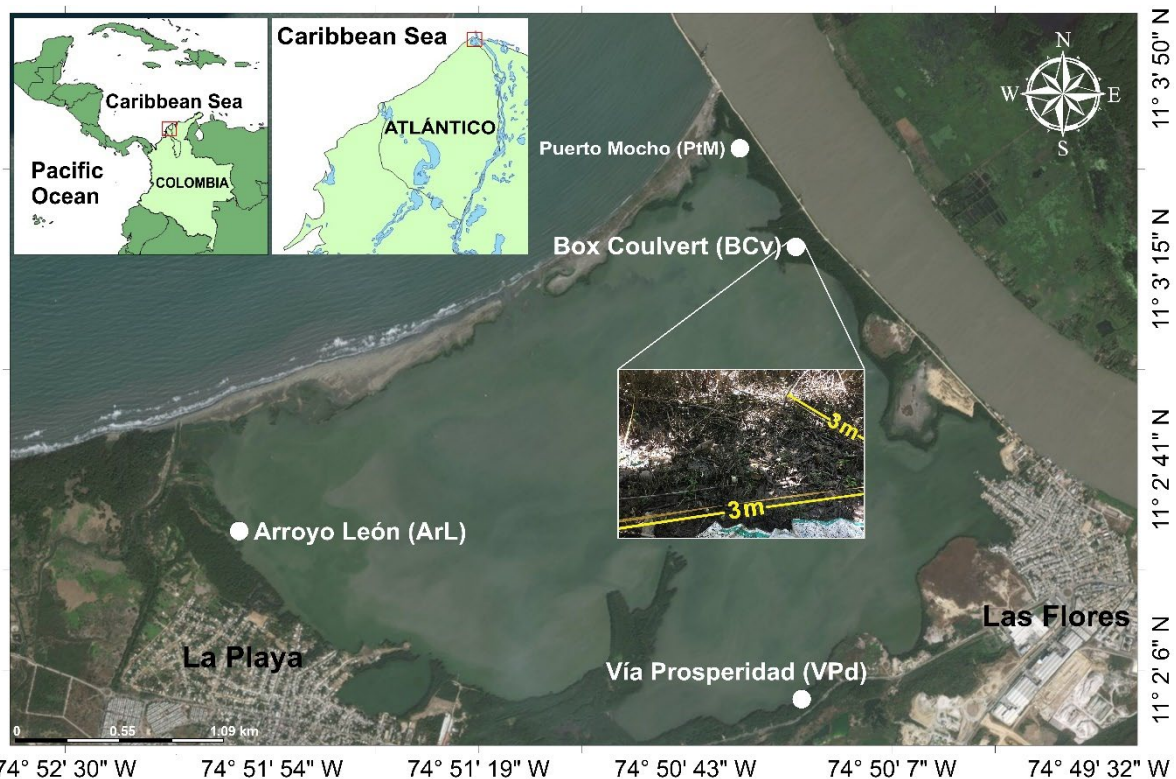


Figure 1. Location of the CM and surveyed sites inside the mangrove forest with a 9 m<sup>2</sup> quadrat in each of the sample points.

The CM is part of the 225,000-ha northern floodplain of the Magdalena River. It communicates with the Caribbean Sea sporadically by natural and human-made events, with one or more openings in the sand bar (Fig. 1). Salinity ranges between 9 and 22 PSU in surface waters and between 14 and 50 PSU in interstitial waters (Chacón *et al.*, 2020; Villate *et al.*, 2020). The CM receives water contributions mainly from hydrographic basins such as Arroyo León and Arroyo Grande—intermittent channels that only provide large volumes of water during the rainy season. The main channel is Arroyo León, with a length of 37 km and a hydrographic area of 247 km<sup>2</sup> (Vera and De La Rosa Muñoz, 2003; GTA, 2005; Fuentes *et al.*, 2018).

The CM has three types of vegetation: 1) herbaceous vegetation such as bushes and paddocks; 2) mangroves (*Rhizophora mangle* L., *Avicennia germinans* (L.) Stearn, *Laguncularia racemosa* (L.) C. F. Gaertn. and eventually *Conocarpus erectus*) and 3) scrub (Vera and De La Rosa Muñoz, 2003).

Mangrove forests make the main structure covering the CM ecosystem, with mainly mixed formations of the basin physiographic type, developed in areas with slow water exchange and high sedimentation rates. There are also monospecific and mixed formations made of three or four species, dominated by *A. germinans*.

The study was conducted at four sites: (1) Puerto Mocho site (PtM) is a forest made up of two mangrove species (1660 trees ha<sup>-1</sup>, 20.2 m<sup>2</sup>·ha<sup>-1</sup>), dominated by *R. mangle*. (2) Box Couvert site (BCv) is a mixed mangrove forest composed of *A.*

*germinans*, *L. racemosa* and *R. mangle* (1360 trees ha<sup>-1</sup>, 23.6 m<sup>2</sup>·ha<sup>-1</sup>). (3) Vía Prosperidad site (VPd) is a narrow strip of mixed mangrove forest composed of *L. racemosa*, *R. mangle*, and *A. germinans*, with good structural development, high conservation status, and excellent phytosanitary conditions. (4) Arroyo León site (ArL) is a mixed forest (2016.7 trees ha<sup>-1</sup>, 15.8 m<sup>2</sup>·ha<sup>-1</sup>) composed of *L. racemosa*, *R. mangle*, and *A. germinans*. ArL is the site with the highest density and the lowest total basal area, suggesting a lower degree of development or a greater degree of degradation.

Despite the degradation of the mangrove forest, the ecosystem has a high diversity of species such as snails, oysters, clams, shrimp, and crabs; and fish species like mullet, tarpon, snook, mojarra, anchovies, bonefish, catfish, and others of great commercial and environmental importance. Bird species are still abundant, mainly pelicans, seagulls, and flamingos (Padilla-Barrios and Pineda-Vides, 2019).

## 2.2 Litter collection and quantification

The ML-associated pollution in the mangrove forest of the CM was executed by adapting methodologies and international guidelines for monitoring marine litter developed by the United Nations Environment Program (UNEP, 2009) and the OSPAR Commission (2010). ML was collected in an area of 9 m<sup>2</sup> located close to the body of water in each of the sample points (Fig. 1). The litter density and weight were calculated in terms of the number of items·m<sup>-2</sup> and kg·m<sup>-2</sup>, respectively.

ML types were divided and counted *in situ*. The OSPAR Commission (2010) proposed a guide used for classifying the ML based on properties, origin, and uses. A total of 135 categories were identified and then grouped into general categories such as plastic, polystyrene, rubber, paper/cardboard, textile, processed wood, glass, metal, medical disposals, fishing gear, organic matter (e.g., pet excrement), cigarette butts, mixed (footwear made of different materials such sandals), and other (unidentified materials).

2.3 ML sources

The quantified ML was classified according to its buoyancy, following the method proposed by Rech et al. (2014) into:

1. **Persistent buoyancy objects:** litter items that can potentially float without sinking or decomposing (e.g., plastic, polystyrene, and processed wood).
2. **Intermediate buoyancy objects:** short-term litter items transported by currents that sink or decompose in a relatively short time (e.g., rubber, textiles, paper/cardboard, fishing gear, mixed, and organic matter).
3. **Non-floating objects:** dense or heavy litter items transported for short distances from the source (e.g., metal and glass).

The Ocean Conservancy proposed a method (2010) to relate the ML items to their sources (economic sector or human activity). The categories of possible sources of marine litter are: 1) shoreline and recreational activities; 2) marine and river activities; 3) smoking-related activities; 4) dumping activities, and 5) disposal of personal care and hygiene items.

2.4 Environmental impact of ML

The environmental impact of ML in the mangrove forest of the CM was assessed employing two indexes: The Clean-Coast Index (CCI) modified in this study going from using the total number of articles to their density (not overestimating the results due to the small sample area used on each point), and the Hazardous Items Index (HII). The modified CCI was calculated by quantifying the ML in each mangrove forest study site, based on the index proposed by Alkalay et al. (2007), using the following equation:

$$CCI = \frac{\Sigma Total\ items/m^2}{Area} * K$$

where CCI determines the state of cleanliness (absence of ML) and the habits required to achieve and maintain that state (Rangel-Buitrago et al., 2019a, 2019b), estimated by total items quantified, divided by the total sampling area (m<sup>2</sup>) and multiplied by the K coefficient equal to 20. Based on the scale provided by Alkalay et al. (2007), resulting values are grouped into five different classes, ranging from I ‘very clean’ to V ‘extremely dirty’.

The HII proposed by Rangel-Buitrago et al. (2019b), measures the possibility of being affected by hazardous marine litter (HML) items that can generate a potential danger (direct or indirect) for any living organism. These are grouped into sharp litter (e.g., metal, glass, and ceramics) and toxic litter (e.g., medications, health, and sanitary paraphernalia). The amount of

potentially hazardous litter items (both sharp and toxic) is calculated using the following equation:

$$HII = \frac{\frac{\Sigma\ total\ hazardous\ litter\ items}{Log_{10}(\Sigma\ total\ litter\ items)}}{Area} * K$$

HII evaluates the relationship between hazardous litter items and Log<sub>10</sub> of the total number of litter items found in the sampled area. K is a coefficient equal to 8, used to generate a better interpretation of the data. In terms of exposure to hazardous litter, the environmental quality of the ecosystem is classified into five classes that range from I ‘Safe site – no hazardous items’ to V ‘Extremely hazardous site – too many hazardous items.’

2.5 CCI vs HII in the CM

Following the method proposed by Rangel-Buitrago et al. (2019a), the CCI and the HII were integrated through a sector analysis developed by Williams et al. (2016a, 2016b). A table was constructed using the percentile technique (Langford, 2006) and was divided into three areas:

- The green area shows from very clean and clean areas without hazardous items, where protection measures are necessary to maintain current conditions.
- The orange area represents moderate cleanliness sites with a considerable number of hazardous items and where cleaning actions are necessary.
- The red area describes from dirty to extremely dirty areas with a large number of hazardous items where urgent intervention and even restoration measures are necessary.

3. Results

3.1 Marine litter characterization

A total of 860 items were collected in the 36 m<sup>2</sup> area of mangrove forest in the CM; these items weighed 77.9 kg of ML and had an average item density of 23.89 items\*m<sup>-2</sup>, and an average weight density of 2.16 kg\*m<sup>-2</sup> (Table 1, Fig. 2).

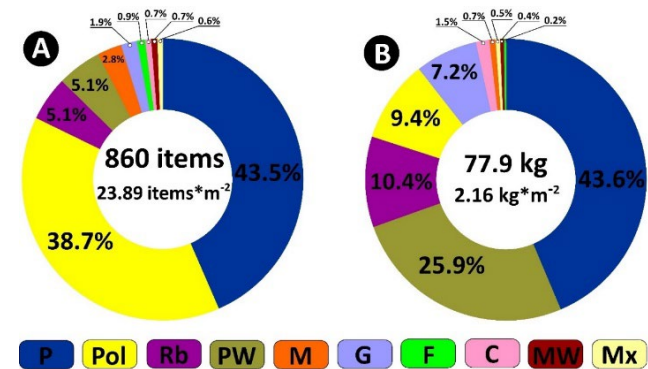


Figure 2. Magnitude and composition of marine litter: (A) number of items and (B) weight (kg). Plastic [P], polystyrene [Pol], rubber [Rb], processed wood [PW], metal [M], glass [G], fishing gear [F], cloth [C], medical disposals [MW], and mixed items [Mx].

The densities of ML varied depending on the sampled sites. The Arroyo León site (ArL) registered the highest density, with 57.11 items\*m<sup>-2</sup> and 60.0 % of the total ML collected in the CM, followed by the Via Prosperidad site (VPd), with a density of 30.56 items\*m<sup>-2</sup> and 32.0 % of the total ML collected (Table 1). Together, Puerto Mocho (PtM) and Box Couvert (BCv) represented 8.0 % of the total ML recorded in our study, with densities of 1.67 items\*m<sup>-2</sup> and 6.22 items\*m<sup>-2</sup>, respectively (Table 1).

Still, plastic and polystyrene were the most abundant litter

items registered, each one representing 43.5 % (10.38 items\*m<sup>-2</sup>) and 38.7 % (9.25 items\*m<sup>-2</sup>) of all ML collected. Which was true in each site sampled, where the plastic composition was between 41.5 % and 66.7 %, and that of polystyrene was between 28.6 % and 40.5 %. Other items found were rubber (5.1%), processed wood (5.1 %), metal (2.8 %), and glass (1.9 %). Textiles, medical disposals, fishing gear, and mixed items such as sandals made up only 0.9 % to 0.6 % of the items collected (Table 1, Fig. 2a).

Table 1.

Categories, types, sources, items number, densities (items\*m<sup>-2</sup>), and percentages of marine litter observed at each site inside the mangrove forest of the Ciénaga de Mallorquín. Sites: Puerto Mocho (PtM), Box Couvert (BCv), Via Prosperidad (VPd), and Arroyo León (ArL). Plastic [P], polystyrene [Pol], rubber [Rb], processed wood [PW], metal [M], glass [G], fishing gear [F], cloth [C], medical disposals [MW], and mixed items [Mx].

Ospar Code	Item	Type	Source	PtM		BCv		VPd		ArL		Total		
				N°	Density	N°	Density	N°	Density	N°	Density	N°	Density	%
2	Bags	[P]	Dumping	0	0.00	0	0.00	26	2.89	0	0.00	26	0.72	3.02
3	Small plastic bags	[P]	Dumping	0	0.00	0	0.00	4	0.44	0	0.00	4	0.11	0.47
4	Drinks (bottles, containers, and drums)	[P]	Shoreline and Recreational	3	0.33	4	0.44	16	1.78	82	9.11	105	2.92	12.21
5	Cleaner (bottles, containers, and drums)	[P]	Dumping	1	0.11	2	0.22	2	0.22	3	0.33	8	0.22	0.93
6	Food containers ind. Fast food containers	[P]	Dumping	0	0.00	1	0.11	5	0.56	2	0.22	8	0.22	0.93
7	Cosmetics	[P]	Medical/Personal Hygiene	1	0.11	4	0.44	2	0.22	12	1.33	19	0.53	2.21
*	Cooking oil containers	[P]	Dumping	0	0.00	4	0.44	10	1.11	16	1.78	30	0.83	3.49
8	Engine oil containers and drums <50 cm	[P]	Ocean/water way	0	0.00	3	0.33	1	0.11	1	0.11	5	0.14	0.58
12	Other bottles, containers and drums	[P]	Dumping	0	0.00	2	0.22	4	0.44	0	0.00	6	0.17	0.70
14	Car parts / Electric appliances	[P]	Dumping	1	0.11	0	0.00	5	0.56	2	0.22	8	0.22	0.93
15	Caps / lids	[P]	Shoreline and Recreational	3	0.33	13	1.44	26	2.89	69	7.67	111	3.08	12.91
16	Cigarette lighters	[P]	Smoking-Related	0	0.00	0	0.00	2	0.22	4	0.44	6	0.17	0.70
17	Pens	[P]	Dumping	0	0.00	0	0.00	1	0.11	2	0.22	3	0.08	0.35
18	Combs / hairbrushes	[P]	Dumping	0	0.00	1	0.11	0	0.00	4	0.44	5	0.14	0.58
*	Toothbrush	[P]	Dumping	0	0.00	0	0.00	2	0.22	1	0.11	3	0.08	0.35
*	Kitchen utensils	[P]	Dumping	0	0.00	0	0.00	0	0.00	1	0.11	1	0.03	0.12
20	Toys and party poppers	[P]	Dumping	0	0.00	0	0.00	3	0.33	6	0.67	9	0.25	1.05
21	Cups	[P]	Shoreline and Recreational	1	0.11	1	0.11	3	0.33	5	0.56	10	0.28	1.16
22	Cutlery / trays / straws	[P]	Shoreline and Recreational	0	0.00	0	0.00	1	0.11	4	0.44	5	0.14	0.58
23	Mesh vegetable bags	[P]	Dumping	0	0.00	0	0.00	1	0.11	1	0.11	2	0.06	0.23
48	Other plastic items	[P]	Dumping	0	0.00	0	0.00	0	0.00	19	2.11	19	0.53	2.21
6	Food containers	[Pol]	Dumping	0	0.00	1	0.11	7	0.78	5	0.56	13	0.36	1.51

Ospar Code	Item	Type	Source	PtM		BCv		VPd		ArL		Total		
				N°	Density	N°	Density	N°	Density	N°	Density	N°	Density	%
	ind. Fast food containers													
45	Foam sponge	[Pol]	Ocean/water way	0	0.00	0	0.00	8	0.89	18	2.00	26	0.72	3.02
117	Polystyrene pieces 0-2,5 cm	[Pol]	Dumping	0	0.00	0	0.00	16	1.78	46	5.11	62	1.72	7.21
46	Polystyrene pieces 2,5-50 cm	[Pol]	Dumping	1	0.11	15	1.67	61	6.78	110	12.22	187	5.19	21.74
47	Polystyrene pieces >50 cm	[Pol]	Dumping	0	0.00	0	0.00	16	1.78	10	1.11	26	0.72	3.02
49	Ballons, valves, ribbons, strings, etc.	[Rb]	Dumping	0	0.00	0	0.00	1	0.11	1	0.11	2	0.06	0.23
50	Boots/Shoes/sandals	[Rb]	Shoreline and Recreational	1	0.11	0	0.00	4	0.44	34	3.78	39	1.08	4.53
52	Tires and belts	[Rb]	Dumping	0	0.00	0	0.00	0	0.00	1	0.11	1	0.03	0.12
53	Other rubber pieces	[Rb]	Dumping	0	0.00	0	0.00	1	0.11	1	0.11	2	0.06	0.23
57	Shoes (e.g., leather)	[C]	Dumping	0	0.00	0	0.00	1	0.11	4	0.44	5	0.14	0.58
59	Other textiles	[C]	Dumping	0	0.00	0	0.00	0	0.00	1	0.11	1	0.03	0.12
74	Other woods <50 cm	[PW]	Dumping	1	0.11	2	0.22	4	0.44	4	0.44	11	0.31	1.28
75	Other woods >50 cm	[PW]	Dumping	0	0.00	0	0.00	13	1.44	20	2.22	33	0.92	3.84
77	Bottle caps	[M]**	Dumping	0	0.00	0	0.00	2	0.22	6	0.67	8	0.22	0.93
89	Other metal pieces	[M]**	Dumping	0	0.00	0	0.00	13	1.44	3	0.33	16	0.44	1.86
91	Bottles	[G]**	Dumping	0	0.00	1	0.11	3	0.33	8	0.89	12	0.33	1.40
92	Light bulbs / tubes	[G]**	Dumping	0	0.00	0	0.00	1	0.11	0	0.00	1	0.03	0.12
*	Cosmetic containers (e.g., perfumes)	[G]**	Medical/Personal Hygiene	1	0.11	0	0.00	1	0.11	1	0.11	3	0.08	0.35
103	Medicine containers	[MW]**	Medical/Personal Hygiene	1	0.11	2	0.22	1	0.11	2	0.22	6	0.17	0.70
115	Nets and pieces of net < 50 cm	[F]	Ocean/water way	0	0.00	0	0.00	1	0.11	0	0.00	1	0.03	0.12
33	Fragments fishing nets	[F]	Ocean/water way	0	0.00	0	0.00	6	0.67	1	0.11	7	0.19	0.81
*	Footwear	[Mx]	Dumping	0	0.00	0	0.00	1	0.11	4	0.44	5	0.14	0.58
	<b>Total</b>			<b>15</b>	<b>1.67</b>	<b>56</b>	<b>6.22</b>	<b>275</b>	<b>30.56</b>	<b>514</b>	<b>57.11</b>	<b>860</b>	<b>23.89</b>	<b>100.00</b>
	<b>HML</b>			<b>2</b>	<b>0.22</b>	<b>3</b>	<b>0.33</b>	<b>21</b>	<b>2.33</b>	<b>20</b>	<b>2.22</b>	<b>46</b>	<b>1.28</b>	<b>5.35</b>
	<b>CCI</b>			<b>371</b>	<b>-</b>	<b>1382</b>	<b>-</b>	<b>6791</b>	<b>-</b>	<b>12691</b>	<b>-</b>	<b>1327</b>	<b>-</b>	<b>-</b>
	<b>HII</b>			<b>151</b>	<b>-</b>	<b>152</b>	<b>-</b>	<b>765</b>	<b>-</b>	<b>656</b>	<b>-</b>	<b>348</b>	<b>-</b>	<b>-</b>

\* Litter items not described by [Ocean Conservancy \(2010\)](#).

\*\* HML collected in this study.

In terms of weight, plastic (43.6 %, 0.94 kg\*m<sup>-2</sup>) and processed wood (25.9 %, 0.56 kg\*m<sup>-2</sup>) were the predominant items in the litter collected due to their large volume and higher density. Next was rubber (0.23 kg\*m<sup>-2</sup>), polystyrene (0.20 kg\*m<sup>-2</sup>), and glass (0.16 kg\*m<sup>-2</sup>) (Fig. 2b). This result, both in terms of number of items and weight, highlights the fact that plastics were the main ML items collected in the mangrove forest of the CM, with a total of 374 items and 34 kg (Table 1, Fig. 2).

Table 2.

Buoyancy (%) of ML at each site and along the study area surveyed in the Ciénaga de Mallorquín. Sites: Puerto Mocho (PtM), Box Coulvert (BCv), Vía Prosperidad (VPd), and Arroyo León (ArL).

Buoyancy	PtM	BCv	VPd	ArL	TOTAL
Persistent	80.0	94.6	86.9	87.0	87.3
Intermediate	13.3	3.6	5.8	9.5	8.0
No buoyancy	6.7	1.8	7.3	3.5	4.7



Figure 3. Different kinds of ML were collected in the mangrove forest of the Ciénaga de Mallorquín. Items: plastic (A), glass containers (B), rubber footwear (C), food packaging (D), remains of fishing gear (E), and medicine containers (F).

### 3.2 Buoyancy and sources of ML

The 87.3% of items collected in the study area were of persistent buoyancy, followed by intermediate buoyancy (8.0 %) and non-buoyancy (4.7 %). Similar results were registered in the remaining sample sites, with ML of persistent buoyancy between 80.0 % and 94.6 % (Table 2).

Based on the [Ocean Conservancy](#) method (2010), 43 categories of litter were identified, consisting of plastic items (21), polystyrene (5) (Fig. 3a, d), rubber (4) (Fig. 3c), glass (3) (Fig. 3f), processed wood (2), metal (2), fishing gear (2) (Fig. E), textile (2), medical disposals (1) (Fig. 3b) and varied items (1). Most items collected come from dumping activities at 59.2 % (14.14 items\*m<sup>-2</sup>) (e.g., cooking oil containers and toys), followed by the shoreline and recreational activities with 31.4 % (7.50 items\*m<sup>-2</sup>) (e.g., soda containers, caps/lids, and footwear), ocean/waterway with 4.5 % (e.g., fishing gear leftovers), personal care and hygiene with 4.2 % (e.g., cosmetic and medicine containers) and smoking-related with

0.7 % (cigarette lighters) (Table 3, Fig. 3).

1. Mainly littering by beach tourists, but also from waterside sports or debris washed away from streets, drains, and gutters.
2. Solid waste from recreational fishing/boating, recreational/commercial fishing, recreational/commercial shipping, and the oil and gas industry.
3. In Colombia, cigarette lighters are commonly used for cooking activities.
4. ML is derived from building and construction materials, tires, car parts, drums, household trash, and appliances.
5. Materials discarded into toilets/sewer system or left behind by beach tourists.

The BCv, VPd, and ArL sites had mainly litter from dumping activities, given their proximity to the mouth of tributaries of the Arroyo León hydrographic basin and the Magdalena River basin (Fig. 4a, b) and also their proximity to open-air Sanitary landfills.

Table 3.

The number of items, percentages, and average density (items\*m<sup>-2</sup>) of ML in the mangrove forest of the Ciénaga de Mallorquín concerning sources and activities. Sites: Puerto Mocho (PtM), Box Couvert (BCv), Vía Prosperidad (VPd), and Arroyo León (ArL).

Source/Activity	PtM		BCv		VPd		ArL		Total		
	N°	%	N°	%	N°	%	N°	%	N°	Avg. Density	%
Shoreline and Recreational (1)	8	53.33	18	32.14	50	18.18	194	37.74	270	7.50	31.40
Ocean/waterway (2)	0	0.00	3	5.36	16	5.82	20	3.89	39	1.08	4.53
Smoking-Related (3)	0	0.00	0	0.00	2	0.73	4	0.78	6	0.17	0.70
Dumping (4)	4	26.67	28	50.00	201	73.09	276	53.70	509	14.14	59.19
personal care and hygiene (5)	3	20.00	7	12.50	6	2.18	20	3.89	36	1.00	4.19
<b>Total</b>	<b>15</b>	<b>100</b>	<b>56</b>	<b>100</b>	<b>275</b>	<b>100</b>	<b>514</b>	<b>100</b>	<b>860</b>	<b>23.89</b>	<b>100</b>



Figure 4. Evidence of marine litter from the estuaries of a) the Arroyo León Stream near site ArL, b) the Magdalena River near site BCv, and c) litter dumps near the mangrove forest.

The PtM site located northeast of the CM and far from stream mouths—registered a higher composition of litter accumulated at the shoreline and from recreational activities (Table 3). The PtM site is located also, near a residential area on the Puerto Mocho beach with a population that occasionally dumps its trash outdoors on the margins of the mangrove lagoon (Fig. 4c, d).

### 3.3 CCI vs HII in the CM

The CM with 860 articles of marine litter obtained a dirty state with a value of 13.27 based on the CCI (Class IV). In turn, the CM presented a considerable quantity of HML with a value of 3.48 based on the HII (Class III) for the 46 hazardous litter items that were found (5.3 % of the total collected throughout the study area) and classified as either sharp (glass and metal) or toxic (medicine containers) (Table 1, Fig. 2a). The CM obtained this status due to the entry and accumulation of ML in three of the four evaluated sites, with the VPd and ArL classified as extremely dirty based in the CCI (Class V, with densities of 30.56 items\*m<sup>-2</sup> and 57.11 items\*m<sup>-2</sup>) and was a lot HML based in the HII (Class IV, with densities of 2.33 items\*m<sup>-2</sup> and 2.22 items\*m<sup>-2</sup>), respectively. The BCv site was classified as dirty based in the CCI (Class IV, with a density of 6.22 items\*m<sup>-2</sup>) and had a considerable amount of HML based in the HII (Class III, with a density of 0.33 items\*m<sup>-2</sup>). The PtM was the only site considered clean based in the CCI (Class I, with a density of 1.67 items\*m<sup>-2</sup>) and presented a considerable amount of HML based in the HII (Class III, with a density of 0.22 items\*m<sup>-2</sup>) (Table 1, Fig. 5).

SECTOR ANALYSIS		HAZARDOUS ITEMS INDEX (HII)					
		I	II	III	IV	V	
CLEAN-COAST INDEX (CCI)	Very clean	0	0	0	0	0	0
	Clean	0	0	1	0	0	1
	Moderate	0	0	0	0	0	0
	Dirty	0	0	1	0	0	1
	Extremely dirty	0	0	0	2	0	2
		0	0	2	2	0	4

PROTECTION      CLEANING      RESTORATION

Figure 5. Sector Analysis Approach. Integration of the Clean Coast Index (CCI) with the Hazardous Items Index (HII).



This type of litter consisted of sharp objects, such as metal and glass (40 items, see Fig. 3f), or toxic items like medicine containers (6 items, see Fig. 3b), that come from dumping activities (37 items), personal care and hygiene debris (9 items) (Table 1).

#### 4. Discussion

The 23.89 items\*m<sup>-2</sup> in the CM represent a large amount of litter from different sources and litter-generating activities and make it one of the mangrove ecosystems with the highest density of ML among different studies published around the world. Similar results were recorded by Smith (2012) in Papua New Guinea, Australia (15.29 items\*m<sup>-2</sup>); Fernandino *et al.* (2016) in the Estuarine System of Santos - São Vicente, Brazil (7.80 items\*m<sup>-2</sup>); Riascos *et al.* (2019) in the Buenaventura Bay, Colombia (13.25 items\*m<sup>-2</sup>); and Chee *et al.* (2020) in the Mangroves on Penang Island, Malaysia (8.15 items\*m<sup>-2</sup>) all representing high values of ML density (Table 4).

The density of 1.28 items\*m<sup>-2</sup> of HML registered in this study is higher than the reported in other studies of sites with large volumes of ML with a considerable amount of sharp and toxic litter items. For example, Cordeiro and Costa (2010) registered a density of 0.14 items\*m<sup>-2</sup>, Fernandino *et al.* (2016) 0.07 items\*m<sup>-2</sup>, Krelling *et al.* (2017) 0.03 items\*m<sup>-2</sup>, Martín *et al.* (2019) 0.08 items\*m<sup>-2</sup>, Mazarrasa *et al.* (2019) 0.32 items\*m<sup>-2</sup>, Chee *et al.* (2020) 0.47 items\*m<sup>-2</sup>, and Gonçalves *et al.* (2020) 0.18 items\*m<sup>-2</sup>. Currently, the highest density of HML for mangroves was registered by Riascos *et al.* (2019) in Buenaventura Bay, Colombia, with a density of 2.81 items\*m<sup>-2</sup> (Table 4).

The average ML density in weight in the mangrove forest stands of the CM was 2.16 kg\*m<sup>-2</sup>, confirming that this is an

ecosystem with a high input of litter. For example, on the 28<sup>th</sup> of October, 2017, the Secretariat of the District of Urban Control and Public Space of Barranquilla, the environmental authority 'Barranquilla Verde', and volunteers from the civil society and fishermen's associations collected 5 tons of solid litter in 2200 m<sup>2</sup> of mangrove inside the CM; these 5 tons represent an average density in weight of 2.27 kg\*m<sup>-2</sup>. At the same time, on the 6<sup>th</sup> of February, 2021, as part of a Neotropical Census of Waterbirds; and clean-up activities carried out in the CM, 84 kg of recyclable and 62 kg of non-recyclable litter were collected. These values exceed by up to an order of magnitude those registered by Cordeiro and Costa (2010) in the São Vicente Estuary, Brazil (0.13 kg\*m<sup>-2</sup>) and Riascos *et al.* (2019) in the Buenaventura Bay, Colombia (0.002 – 0.31 kg\*m<sup>-2</sup>), becoming the ecosystems with the highest values of ML by weight per square meter of mangrove.

Plastic and polystyrene (another kind of plastic, but separated in the typology due to its high magnitude in the study area), with densities of 10.39 items\*m<sup>-2</sup> and 9.25 items\*m<sup>-2</sup>, respectively, are the common denominator in ML in the mangrove forest of the CM (82.2 %) (Table 1, Fig. 3a), a condition common to the Colombian Caribbean coast (Williams *et al.* 2016a, 2016b, Botero *et al.* 2020, Garcia *et al.* 2018, Rangel-Buitrago *et al.* 2017, 2018, 2019a, 2020, Garcés-Ordóñez *et al.* 2020), the mangroves in the Colombian Caribbean coast (Garcés-Ordóñez *et al.*, 2019) and the Colombian Pacific coast (Riascos *et al.*, 2019), as well as all mangrove ecosystems around the world (Cordeiro and Costa, 2010; Fernandino *et al.*, 2016; Krelling *et al.*, 2017; Martín *et al.*, 2019; Mazarrasa *et al.*, 2019; Suyadi and Manullang, 2020) (Table 4).

Table 4. Density (items\*m<sup>-2</sup>) of ML and HML, and composition (%) of plastic (P) in mangrove forests around the world.

Country	Specific zone	Sites	Total area (m <sup>2</sup> )	Average density	Average density HML	[P] (%)	Reference
Brazil	São Paulo Vicente	8	1,600	1.33	0.14	0.62	Cordeiro y Costa 2010
United States	Sal marshes of northern California	15	63,096	0.14	0.02	0.40	Viehman <i>et al.</i> 2011
Papua New Guinea	Bootles Bay	20	219	15.29	0.52	0.90	Smith 2012
Brazil	Paranaguá Estuarine Complex, Paraná state	6	124,518	0.0023	0.0003	0.84	Possatto <i>et al.</i> 2015
Brazil	Estuarine system of Santos – São Vicente	6	300	7.80	0.07	0.90	Fernandino <i>et al.</i> 2016
Brazil	Paranaguá Estuarine Complex, Paraná state	6	4,500	0.21	0.03	0.58	Krelling <i>et al.</i> 2017
Philippines	Mayo Bay	2	2,400	0.03	0.00	0.95	Abreo <i>et al.</i> 2018
Colombia	Ciénaga Grande de Santa Marta (CGSM)	6	44,445	0.054	0.003	0.94	Garcés-Ordóñez <i>et al.</i> 2019
Colombia	Buenaventura Bay	4	154	13.25	2.81	0.65	Riascos <i>et al.</i> 2019
Saudi Arabia	Red Sea and Arabian Gulf	20	2,263.5	0.66	0.08	0.82	Martín <i>et al.</i> 2019
Spain	Pas, Miera and Asón estuaries	32	1,900	1.53	0.32	0.76	Mazarrasa <i>et al.</i> 2019
Malaysia	Mangroves on Penang Island	3	1,200	8.15	0.47	0.90	Chee <i>et al.</i> 2020
Brazil	Estuary of the Pará River	8	20,286	0.59	0.18	0.15	Gonçalves <i>et al.</i> 2020
Ecuador	Commune Puerto Roma, Province of Guayas	4	70	2.90	0.11	0.71	Jacho 2020
Mauritius	Mangrove forests of Mauritius	2	1,500	1.42	-	0.43	Seeruttun <i>et al.</i> 2021
Colombia	Ciénaga de Mallorquín	4	36	23.89	1.28	0.82	This study

ML pollution of the mangrove forest stands of the CM is aggravated by the large population of the District of Barranquilla (1,239,804 inhabitants), with the highest national litter production levels: about 1.20 kg of litter per person per day, compared to the overall national average 0.90 kg (SSPD, 2017; DANE, 2020). Between 2016 and 2018, Barranquilla generated an average of 46,519.9 tons of litter per month, of which only 2.05 % (954.3 tons) was recycled (SSPD, 2019). Inadequate handling of its solid debris is evident, with part of the litter deliberately thrown into local streams in Barranquilla (Rebolledo-Colina and León-Luna, 2017; Rangel-Buitrago *et al.*, 2020). Despite all the legal framework and all the institutions in charge of issuing the guidelines to regulate litter management (CCO, 2018), Colombia is a country where 11.0 % of its municipalities and cities use open-air dumps to dispose of their litter or dump it directly in the nearest bodies of water (Vivas-Aguas *et al.*, 2015).

Our results confirm other studies carried out in mangrove forests in Colombia. Garcés-Ordóñez *et al.* (2019) and Riascos *et al.* (2019) conclude that solid litter is by far poorly managed, with approximately 65.0 % of coastal municipalities inadequately disposing of their litter causing it to accumulate in mangroves near urban areas. This problem is leading the mangrove forest of the CM to become a litter dump.

The large volumes of ML registered are due to several causes, including the high buoyancy of the litter (Rech *et al.*, 2014, 2018) as observed in this study (Table 2); its resistance and durability (Derraik, 2002; Hardesty *et al.*, 2016; Buhl-Mortensen and Buhl-Mortensen, 2017), the large amount of litter transported from hydrographic basins (Cordeiro and Costa, 2010; Riascos *et al.*, 2019), and the potential of mangroves to retain and accumulate litter (Ivar do Sul *et al.*, 2014; Smith and Edgar, 2014; Martín *et al.* 2019).

Most litter accumulated in the CM comes mainly from continental sources and is transported by the Magdalena River and multiple streams in the basin or deliberately dumped near the ecosystem (Table 3). Similar patterns were found in mangrove areas of Guatemala (Boix-Morán, 2012), on the Colombian Caribbean coast (Rangel-Buitrago *et al.*, 2017, 2020), and in different studies around the world (Cheshire *et al.*, 2009; Ivar do Sul and Costa, 2013; Galgani *et al.*, 2015; Jambeck *et al.*, 2015) where 80 to 90% of ML comes from continental sources, and less than 10% from open sea activities. In the mangrove forest of the CM, dumping activity was determined as the primary source of ML in the ecosystem.

The Arroyo León Basin (where the ArL site is located) and the Arroyo Grande Basin together host the municipalities of Puerto Colombia, Galapa, Baranoa, Tubará, and part of the urban area of the Barranquilla District, totaling a population of 1,386,688 - 53.3 % of the total population of the department (DANE, 2020) and contributing with much of the litter accumulated in the CM. On the other hand, the VPd site piles up ML that comes from the neighborhoods of Las Flores and La Playa, with 78,000 inhabitants (Corredor, 2018). For many

years, Las Flores had a sanitary landfill that flowed directly into the water body of the CM (GTA, 2005), and La Playa lacks an environmental sanitation plan (Rebolledo-Colina and León-Luna, 2017; Berrocal *et al.*, 2018).

The Magdalena River also carries a large volume of ML to the mangrove forest in the CM. The BCv site, with a moderate density of litter, receives its input from litter transported by the Magdalena River (50.0 %) from dumping activities and from the shoreline and recreational activities (32.1 %) (Table 3). The Magdalena River is the prime hydrographic basin in the country, crossing 724 municipalities; unfortunately, 46.0 % of these municipalities do not have an appropriate litter disposal system (Rangel-Buitrago *et al.*, 2017).

Another source of ML in this ecosystem is debris deliberately thrown near the mangrove forest by residents of Puerto Mocho Beach (Fig. 4c). This situation is of concern because, during natural flooding events, this solid litter is carried into the mangrove forest and retained by the aerial roots of the mangrove species (Cordeiro and Costa, 2010; Garcés-Ordóñez and Arenas, 2019; Martín *et al.*, 2019), together with the potential retainer of the plastic of the other typologies of ML (Vélez-Mendoza, 2022).

The high population density in the city of Barranquilla has led to human settlements in mangrove areas (Las Flores and La Playa). The large volumes of ML input from different sources and the inadequate management of natural resources and solid litter have led to ecosystem degradation. The mangrove forest of the CM was classified as having a *dirty state*, with a value close to those of other coastal sites, like the beaches of the Department of Atlántico and Isla Arena in Bolívar, where most locations were classified as dirty or extremely dirty (Rangel-Buitrago *et al.*, 2019a, 2020).

The high density of ML in mangrove stands has collateral effects on the ecosystem and its biota and is also a potential risk for the health and physical integrity of inhabitants and visitors (Somerville *et al.*, 2003; Garcés-Ordóñez and Bayona, 2019; Rangel-Buitrago *et al.*, 2019a, 2019b). Initiatives led by government entities, fishermen's associations, and volunteers to remove litter and plant mangroves in the CM may potentially expose participants to the risk of sharp and toxic litter items such as metal, glass, and medicine containers.

These results joined with those obtained by the sector analysis (Fig. 5), show that large amounts of ML increase the possibility of finding HML harmful to the ecosystem, biota, health, and human physical integrity, highlighting the need for urgent intervention and restoration measures at the VPd and ArL sites. For this reason, visiting these areas is not recommended unless minimum measures of personal care are taken (e.g., use of boots, jeans, and gloves). The PtM and BCv sites receive less ML that still needs adequate management and cleaning actions. Regardless of the condition and status of each site sampled, the mangrove forest of the CM needs the cooperation of all possible actors involved actors: citizens, public and government entities, private sector, and international agreements, together with

efficient, comprehensive, and proactive strategies for controlling and reducing dumping, recycling or reusing waste, and eliminating ML sources.

The Development Bank of Latin America (*Corporación Andina de Fomento*, CAF) is currently designing, formulating, and financing an ambitious recovery project for the CM "focused on the environmental management of the Ciénaga de Mallorquín, to promote low-carbon and climate-resilient development." The project includes an entire urban development articulated with the natural landscape through "stilt structures above the ecosystem, with fishing docks, and areas for recreational activities and water sports, and especially a space for interaction with nature." However, the project does not mention a plan for integral management of the ecosystem.

Our results demonstrate the urgent need to design and implement actions aimed at the integral recovery of the ecosystem. Therefore, the previous project may be the best opportunity to formulate and develop an Integrated Coastal Marine Management plan, where different sectors like the economic, productive, environmental, social, authorities, environmental and territorial entities, and local communities could participate in the recovery and protect this essential ecosystem, referred to as 'Barranquilla's lung' and turn it into 'Barranquilla's biodiversity' showcase.

## 5. Conclusions

Despite the various and recurring cleanup activities carried out in recent years in the CM mangrove, this ecosystem continues to suffer from a high input of ML. Among the primary sources are the litter discharge activities from the urban area and transported to the ecosystem by hydrographic basins (eg, ArL site), and that caused by the defective or non-existent management of litter generated in recreational and shoreline activities in urban centers located near to the ecosystem (eg, VPd site).

The high average density of ML in the CM confirms that its mangrove forest presents a high influx of ML, causing three of the four sites evaluated to require urgent restoration measures, mainly due to the dominance of plastic items. Plastic, due to its persistent buoyancy and ability to retain other types of litter along with the mangrove roots is negatively influencing the cleanliness and safety of the ecosystem, its biota, and the health and physical integrity of the people (eg, residents of the area who depend on the natural resources provided by the ecosystem) due to the presence and accumulation of HML.

## Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

## Acknowledgments

This work was carried out with data obtained during

professional practice in the framework of the inter-administrative agreement 002 signed with Establecimiento Público Ambiental Barranquilla Verde and the University of Atlántico. The authors thank Luis Avila and all of the members of the Fishermen's Association of La Playa township for their constant guidance and collaboration in the field. Also, our acknowledgments to Valeria Barraza and Niny Suarez, for contributing information and collaborating in the field.

## References

- Alkalay, R., Pasternak, G., and Zask, A. (2007). Clean-coast index- A new approach for beach cleanliness assessment. *Ocean and Coastal Management*, 50(5–6), pp. 352–362. DOI: <https://doi.org/10.1016/j.ocecoaman.2006.10.002>
- Antão-Barboza, L., Vethaak, A., Lavorante, B., Lundebye, A., and Guilhermino, L. (2018). Marine microplastic debris: An emerging issue for food security, food safety and human health. *Marine Pollution Bulletin*, 133, pp. 336–348. DOI: <https://doi.org/10.1016/j.marpolbul.2018.05.047>
- Blanchar, S., Yoset, W., Polo, M. y Angel, F. (2020). Construcción vivero de mangles y aulas ambientales – Ciénaga Mallorquín barrio las Flores – Barranquilla. Tesis de grado. Universidad Simón Bolívar, Barranquilla, Colombia. Disponible en: <https://hdl.handle.net/20.500.12442/6205>
- Berrocal, J.C., Ortega, A.M., Reales, R.J., González, S. y Calderón, R. (2018). Capítulo 8: Contaminación en la Ciénaga de Mallorquín: una perspectiva sociojurídica. En: Filut, D., Albor-Chadid, L.I., Selec, A., et al. Educación socioambiental. Albor-Chadid, L.I., Ed., Ediciones Universidad Simón Bolívar. Disponible en: <http://hdl.handle.net/20.500.12442/2606>
- Boix-Morán, L. (2012). Impact of the presence of solid waste in mangrove areas in Las Lisas Chiquimulilla, Santa Rosa, Guatemala. Available at: <http://vinculando.org/ecologia/impacto-de-la-presencia-de-desechos-solidos-en-las-zonas-de-manglar.html>
- Botero, C.M., Zielinski, S., Pereira, C.I., León, J.A., Dueñas, L.F., and Puentes, V. (2020). The first report of deep-sea waste in the South-Western Caribbean Sea. *Marine Pollution Bulletin*, 157, art. 111327. DOI: <https://doi.org/10.1016/j.marpolbul.2020.111327>
- Botterell, Z., Beaumont, N., Dorrington, T., Steinke, M., Thompson, R., and Lindeque, P. (2019). Bioavailability and effects of microplastics on marine zooplankton: a review. *Environmental Pollution*, 245, pp. 98–110. DOI: <https://doi.org/10.1016/j.envpol.2018.10.065>
- Buhl-Mortensen, L., and Buhl-Mortensen, P. (2017). Marine waste in the Nordic Seas: distribution composition and abundance. *Marine Pollution Bulletin*, 125(1-2), pp. 260–270. DOI: <https://doi.org/10.1016/j.marpolbul.2017.08.048>
- Bulow, E.S. y Ferdinand, T.J. (2013). El efecto de la basura en la dinámica de los ecosistemas de manglar: un análisis comparativo. Centro de Incidencia Ambiental – CIAM. Smithsonian Tropical Research Institute. McGill University. Disponible en: [https://www.mcgill.ca/pfss/files/pfss/el\\_efecto\\_de\\_la\\_basura\\_en\\_la\\_dinamica\\_de\\_los\\_ecosistemas\\_de\\_manglar\\_un\\_analisis\\_comparativo.pdf](https://www.mcgill.ca/pfss/files/pfss/el_efecto_de_la_basura_en_la_dinamica_de_los_ecosistemas_de_manglar_un_analisis_comparativo.pdf)

- CCO. Comisión Colombiana del Océano. (2018). Política nacional del océano y de los espacios costeros. 37 P. Disponible en: <http://www.cco.gov.co/docs/publicaciones/pnoec-2018-09.pdf>
- Chacón, S., Serrano, M.C., Bolívar-Anillo, H.J., Villate, D.A., Sánchez, H. y Anfuso, G. (2020). Bosques de manglar del Caribe Norte Colombiano: análisis, evolución y herramientas de gestión. *Revista Latinoamericana de Recursos Naturales*, 16(1), pp. 31–54. DOI: <https://doi.org/10.33154/rlrn.2020.01.04>
- Chee, S.Y., Yee, J.C., Carey, D., Yusup, Y., and Gallenger, J.B. (2020). Anthropogenic marine debris accumulation in mangroves on Penang Island, Malaysia. *Journal of Sustainability Science and Management*, 15(6), pp. 36–60. Available at: [https://jssm.umt.edu.my/wp-content/uploads/sites/51/2020/10/4-\\_15.6.pdf](https://jssm.umt.edu.my/wp-content/uploads/sites/51/2020/10/4-_15.6.pdf)
- Cheshire, A., Adler, A., Barbière, J., and Cohen, Y. (2009). UNEP/IOC Guidelines on survey and monitoring of marine waste. UNEP Regional Seas Reports and Studies (No. 186, IOC Technical Series). Available at: <http://hdl.handle.net/20.500.11822/13604>
- Chong, J. (2005) - Protective values of mangrove and coral ecosystems: a review of methods and evidence. IUCN, Gland, Switzerland. Available at: [https://www.iucn.org/sites/dev/files/import/downloads/p\\_r\\_values\\_mangrove\\_coral\\_ecosystems\\_methods\\_eviden ce.pdf](https://www.iucn.org/sites/dev/files/import/downloads/p_r_values_mangrove_coral_ecosystems_methods_eviden ce.pdf)
- Cordeiro, C.A.M.M., and Costa, T.M. (2010). Evaluation of solid residues removed from a mangrove swamp in the São Vicente Estuary, SP, Brazil. *Marine Pollution Bulletin*, 60(10), pp. 1762–1767. DOI: <https://doi.org/10.1016/j.marpolbul.2010.06.010>
- Corredor, L.Y. (2018). Malecón las Flores. un área de mejoramiento portuario con miras al crecimiento y desarrollo fluvial. Universidad La Gran Colombia. Disponible en: <http://repository.ugc.edu.co/handle/11396/3346>
- DANE. Departamento Administrativo Nacional de Estadística. (2018). Boletín técnico: Cuenta Ambiental y Económica de Flujo de Materiales – Residuos Sólidos 2012 - 2016. Disponible en: <https://www.dane.gov.co>
- DANE. Departamento Administrativo Nacional de Estadística. (2020). Colombia: proyecciones de población municipal por área hasta el 2020 del DANE. Disponible en: <https://www.dane.gov.co>
- Derraik, J.G.B. (2002). The pollution of the marine environment by plastic debris: a review. *Marine Pollution Bulletin*, 44(9), pp. 842–852. DOI: [https://doi.org/10.1016/S0025-326X\(02\)00220-5](https://doi.org/10.1016/S0025-326X(02)00220-5)
- Díaz, J.M. (2011). Una revisión sobre los manglares: características, problemáticas y su marco jurídico. *Ra Ximbai*. 7(3), pp. 355–369. Disponible en: <https://www.redalyc.org/pdf/461/46121063005.pdf>
- Fernandino, G., Elliff, C.I., Frutuoso, G.A., Silva, E.V.N.M. da, Gama, G.S., Sousa, J.H. de O., and Silva, I.R. (2016). Considerations on the effects of tidal regimes in the movement of floating waste in an estuarine environment: Case study of the estuarine system of Santos-São Vicente, Brazil. *Marine Pollution Bulletin*, 110(1), pp. 591–595. DOI: <https://doi.org/10.1016/j.marpolbul.2016.05.080>
- Friess, D.A., Rogers, K., Lovelock, C.E., Krauss, K.W., Hamilton, S.E., Lee, S.Y., Lucas, R., Primavera, J., Rajkaran, A., and Shi, S. (2019). The State of the World's Mangrove Forests: Past, Present, and Future. *Annual Review of Environment and Resources*, 44, pp. 89–115. DOI: <https://doi.org/10.1146/annurev-environ-101718-033302>
- Fuentes, F., Pinedo, J. y Marrugo, J. (2018). Metales pesados en especies ícticas de la ciénaga de Mallorquín, Colombia. *Espacios*, 39(3), art. 1225. Disponible en: <http://hdl.handle.net/11323/1225>
- Galgani, F., Hanke, G., and Maes, T. (2015). Global distribution, composition and abundance of marine waste. In: Bergmann, M., Gutow, L. and Klages, M. *Marine Anthropogenic Waste*, Springer, Berlin, pp. 29–56. DOI: <https://doi.org/10.1007/978-3-319-16510-3>
- Gall, S.C., and Thompson, R.C. (2015). The impact of debris on marine life. *Marine Pollution Bulletin*, 92(1–2), pp. 170–179. DOI: <https://doi.org/10.1016/j.marpolbul.2014.12.041>
- Garcés-Ordóñez, O., Castillo-Olaya, V.A., Granados-Briceño, A.F., Blandón, L.M., and Espinosa, L.F. (2019). Marine waste and microplastic pollution on mangrove soils of the Ciénaga Grande de Santa Marta, Colombian Caribbean. *Marine Pollution Bulletin*, 145(2), pp. 455–462. DOI: <https://doi.org/10.1016/j.marpolbul.2019.06.058>
- Garcés-Ordóñez, O. y Bayona, M.R. (2019). Impactos de la contaminación por basura marina en el ecosistema de manglar de la Ciénaga Grande de Santa Marta, Caribe colombiano. *Revista Ciencias Marinas y Costeras*, 11(2), pp. 145–165. DOI: <https://doi.org/10.15359/revmar.11-2.8>
- Garcés-Ordóñez, O., Espinosa, L.F., Pereira, R., and Costa, M. (2020). The impact of tourism on marine waste pollution on Santa Marta beaches, Colombian Caribbean. *Marine Pollution Bulletin*, 160, art. 111558. DOI: <https://doi.org/10.1016/j.marpolbul.2020.111558>
- Giesen, W., Wulffraat, S., Zieren, M., and Scholten, L. (2007). *Mangrove guidebook for Southeast Asia*. Food and Agricultural Organization and Wetlands International, Bangkok. Available at: <http://www.fao.org/3/ag132e/ag132e00.htm>
- Gonçalves, M., Schmid, K., Andrade, M.C., Andrades, R., Pegado, T., and Guiarizzo, T. (2020). Are the tidal flooded forests sinks for litter in the Amazonian estuary? *Marine Pollution Bulletin*, 161, art. 111732. DOI: <https://doi.org/10.1016/j.marpolbul.2020.111732>
- Gracia, A., Rangel-Buitrago, N., and Flórez, L. (2018). Beach waste and woody-debris colonizers on the Atlántico department Caribbean coastline, Colombia. *Marine Pollution Bulletin*, 128, pp. 185–196. DOI: <https://doi.org/10.1016/j.marpolbul.2018.01.017>
- Green, D.S., Boots, B., Blockley, D.J., Rocha, C., and Thompson, R. (2015). Impacts of discarded plastic bags on marine assemblages and ecosystem functioning. *Environmental Science and Technology*, 49(9), pp. 5380–5389. DOI: <https://doi.org/10.1021/acs.est.5b00277>
- GTA. Grupo de Investigación en Tecnologías del Agua. (2005). Análisis sobre el manejo integral del recurso hídrico en la Ciénaga de Mallorquín. Informe ejecutivo, Grupo de Investigación en Tecnologías en el agua. Universidad del Norte, Barranquilla. Disponible en: <https://www.yumpu.com/xx/document/view/56802314/analisis-sobre-el-manejo-integrado-del-recurso-hidrico-en-la-cienaga-de-mallorquin>

- Hamilton, S.E., and Casey, D. (2016). Creation of a high Spatio-temporal resolution global database of continuous mangrove forest cover for the 21st century (CGFMC-21). *Global Ecology and Biogeography*, 25(6), pp. 729–738. DOI: <https://doi.org/10.1111/geb.12449>
- Hardesty, B.D., Lawson, T.J., van der Velde, T., Lansdell, M., and Wilcox, C. (2016). Estimating quantities and sources of marine debris at a continental scale. *Frontiers and Ecology and the Environment*, 1, pp. 18–25. DOI: <https://doi.org/10.1002/fee.1447>
- Hartley, B.L., Thompson, R.C., and Pahl, S. (2015). Marine waste education boosts children's understanding and self-reported actions. *Marine Pollution Bulletin*, 90(1–2), pp. 209–217. DOI: <https://doi.org/10.1016/j.marpolbul.2014.10.049>
- Ivar do Sul, J.A., and Costa, M.E. (2013). Plastic pollution risks in an estuarine conservation unit. *Journal of Coastal Research*, 65, pp. 48 – 53. DOI: <https://doi.org/10.2112/SI65-009.1>
- Ivar do Sul, J.A., Costa, M.F., Silva-Cavalcanti, J.S., and Araújo, M.C.B. (2014). Plastic debris retention and exportation by a mangrove forest patch. *Marine Pollution Bulletin*, 78(1–2), pp. 252–257. DOI: <https://doi.org/10.1016/j.marpolbul.2013.11.011>
- Jacho, C.E. (2020). Abundancia y distribución de macrobasura en los manglares de la comuna Puerto Roma, Provincia de Guayas, Ecuador. Tesis de pregrado. Universidad de Guayaquil. Disponible en: <http://repositorio.ug.edu.ec/handle/redug/48725>
- Jambeck, J.R., Geyer, R., Wilcox, C., Siegler, T.R., Perryman, M., Andrady, A., Narayan, R., and Lavender Law, K. (2015). Plastic waste inputs from land into the ocean. *Science*, 347, pp. 1655–1734. DOI: [10.1126/science.1260352](https://doi.org/10.1126/science.1260352)
- Kauffman, J.B., Heider, C., Cole, T.G., Dwire, K.A., and Donato, D.C. (2011). Ecosystem carbon stocks of micronesia mangrove forests. *Wetlands*, 31, pp. 343–352. DOI: <https://doi.org/10.1126/science.1260352> DOI: [10.1007/s13157-011-0148-9](https://doi.org/10.1007/s13157-011-0148-9)
- Krelling, A.P., Souza, M.M., Williams, A.T., and Turra, A. (2017). Transboundary movement of marine waste in an estuarine gradient: Evaluating sources and sinks using hydrodynamic modelling and ground truthing estimates. *Marine Pollution Bulletin*, 119(1), pp. 48–63. DOI: <https://doi.org/10.1016/j.marpolbul.2017.03.034>
- Langford, E. (2006). Quartiles in elementary statistics. *Journal of Statistics Education*, 14(3), art. 10589. DOI: <https://doi.org/10.1080/10691898.2006.11910589>
- Lozoya, J.P., Teixeira de Mello, F., Carrizo, D., Weinstein, F., Olivera, Y., Cedrés, F., Pereira, M., and Fossati, M. (2016). Plastics and microplastics on recreational beaches in Punta del Este (Uruguay): unseen critical residents? *Environmental Pollution*, 218, pp. 931–941. DOI: <https://doi.org/10.1016/j.envpol.2016.08.041>
- Martín, C., Almahasheer, H., and Duarte, C.M. (2019). Mangrove forests as traps for marine waste. *Environmental Pollution*, 247, pp. 499–508. DOI: <https://doi.org/10.1016/j.envpol.2019.01.067>
- Mazarrasa, I., Puente, A., Núñez, P., García, A., Abascal, A.J., and Juanes, J.A. (2019). Assessing the risk of marine waste accumulation in estuarine habitats. *Marine Pollution Bulletin*, 144, pp. 117–128. DOI: <https://doi.org/10.1016/j.marpolbul.2019.04.060>
- Norris, B.K., Mullarney, J.C., Bryan, K.R., and Henderson, S.M. (2017). The effect of pneumatophore density of turbulence: A field study in a *Sonneratia*-dominated mangrove forest, Vietnam. *Continental Shelf Research*, 147, pp. 114–127. DOI: <https://doi.org/10.1016/j.crs.2017.06.002>
- Ocean Conservancy. (2010). Trash travels: from our hands to the sea, around the globe, and through time. Washington, DC. Available at: <https://oceanconservancy.org/wp-content/uploads/2017/04/2010-Ocean-Conservancy-ICC-Report.pdf>
- OSPAR Commission. (2010). Guideline for monitoring marine waste on the beaches in the OSPAR Maritime Area. OSPAR Commission, 1, 84. Available at: [https://www.ospar.org/ospar-data/10-02e\\_beachlitter%20guideline\\_english%20only.pdf](https://www.ospar.org/ospar-data/10-02e_beachlitter%20guideline_english%20only.pdf)
- Padilla-Barrios, J. y Pineda-Vides, F. (2019). Análisis de las condiciones para el desarrollo de un programa de ecoturismo comunitario en la Ciénaga de Mallorquín barrio La Playa, Barranquilla – Colombia. Tesis de grado. Universidad de la Costa (CUC), Barranquilla, Colombia. Disponible en: <http://hdl.handle.net/11323/3118>
- Portz, L., Manzolli, R.P., Andrade, C.F.F., Villate-Daza, D.A., Bolívar D.A.B., and Alcántara-Carrió, J. (2020). Assessment of heavy metals pollution (Hg, Cr, Cd, Ni) in the sediments of Mallorquín lagoon - Barranquilla, Colombia. *Journal Coastal Research*, 95(SI), pp. 158-162. DOI: <https://doi.org/10.2112/SI95-031.1>
- Rangel-Buitrago, N., Williams, A., Anfuso, G., Arias, M., and Gracia, C.A. (2017). Magnitudes, sources, and management of beach waste along the Atlántico department coastline, Caribbean coast of Colombia. *Ocean and Coastal Management*, 138, pp. 142–157. DOI: <https://doi.org/10.1016/j.ocecoaman.2017.01.021>
- Rangel-Buitrago, N., Castro-Barros, J.D., Gracia, A., Villadiego, J.D.V., and Williams, A.T. (2018). Waste impacts on beach/dune systems along the Atlántico Department, the Caribbean Coastline of Colombia. *Marine Pollution Bulletin*, 137, pp. 35–44. DOI: <https://doi.org/10.1016/j.marpolbul.2018.10.009>
- Rangel-Buitrago, N., Gracia, A., Vélez-Mendoza, A., Carvajal-Florián, A., Mojica-Martínez, L., and Neal, W.J. (2019a). Where did this refuse come from? Marine anthropogenic waste on a remote island of the Colombian Caribbean Sea. *Marine Pollution Bulletin*, 149, art. 110611. DOI: <https://doi.org/10.1016/j.marpolbul.2019.110611>
- Rangel-Buitrago, N., Vergara-Cortés, H., Barría-Herrera, J., Contreras-López, M., and Agredano, R. (2019b). Marine debris occurrence along Las Salinas beach, Viña Del Mar (Chile): Magnitudes, impacts and management. *Ocean and Coastal Management*, 178, art. 104842. DOI: <https://doi.org/10.1016/j.ocecoaman.2019.104842>
- Rangel-Buitrago, N., Vélez-Mendoza, A., Gracia, A., and Neal, W.J. (2020). The impact of anthropogenic Waste on Colombia's central Caribbean beaches. *Marine Pollution Bulletin*, 152, art. 110909. DOI: <https://doi.org/10.1016/j.marpolbul.2020.110909>
- Rebolledo-Colina, J.I., and León-Luna, I.M. (2017). Design and calculation of modules for the hydrodynamic reestablishment of the Mallorquín Ciénaga. *Ciencia e Ingeniería*. 4(2). Available at: <http://revistas.uniguajira.edu.co/rev/index.php/cei/article>

- /view/78/79
- Rech, S., Macaya-Caquilpán, V., Pantoja, J.F., Rivadeneira, M.M., Jofre Madariaga, D., and Thiel, M. (2014). Rivers as a source of marine waste - A study from the SE Pacific. *Marine Pollution Bulletin*, 82(1–2), pp. 66–75. DOI: <https://doi.org/10.1016/j.marpolbul.2014.03.019>
- Rech, S., Salmina, S., Borrell, Y., and Garcia-Vasquez, E. (2018). Dispersal of alien invasive species on anthropogenic waste from European mariculture areas. *Marine Pollution Bulletin*, 131, pp. 10–16. DOI: <https://doi.org/10.1016/j.marpolbul.2018.03.038>
- Riascos, J.M., Valencia, N., Peña, E.J., and Cantera, J.R. (2019). Inhabiting the technosphere: The encroachment of anthropogenic marine waste in Neotropical mangrove forests and its use as habitat by macrobenthic biota. *Marine Pollution Bulletin*, 142, pp. 559–568. DOI: <https://doi.org/10.1016/j.marpolbul.2019.04.010>
- Seeruttun, L.D., Raghbor, P., and Appadoo, C. (2021). First assessment of anthropogenic marine debris in mangrove forests of Mauritius, a small oceanic island. *Marine Pollution Bulletin*, 164, art. 112019. DOI: <https://doi.org/10.1016/j.marpolbul.2021.112019>
- Smith, S.D.A. (2012). Marine debris: a proximate threat to marine sustainability in Bootless Bay, Papua New Guinea. *Marine Pollution Bulletin*, 64(9), pp. 1880–1883. DOI: <https://doi.org/10.1016/j.marpolbul.2012.06.013>
- Smith, S.D.A., and Edgar, R.J. (2014). Documenting the density of subtidal marine debris across multiple marine and coastal habitats. *PLoS ONE*, 9(4), art. 94593. DOI: <https://doi.org/10.1371/journal.pone.0094593>
- Somerville, S.E., Miller, K.L., and Mair, J.M. (2003). Assessment of the aesthetic quality of a selection of beaches in the Firth of Forth, Scotland. *Marine Pollution Bulletin*, 46, pp. 1184–1190. DOI: [https://doi.org/10.1016/S0025-326X\(03\)00126-7](https://doi.org/10.1016/S0025-326X(03)00126-7)
- SSPD. Superintendencia de Servicios Públicos Domiciliarios. (2017). Informe nacional de aprovechamiento - 2016. SSPD. Disponible en: [https://www.superservicios.gov.co/sites/default/archivos/SSPD%20Publicaciones/Publicaciones/2018/Oct/informenacionaldeaprovechamiento2016\\_dic1920161.pdf](https://www.superservicios.gov.co/sites/default/archivos/SSPD%20Publicaciones/Publicaciones/2018/Oct/informenacionaldeaprovechamiento2016_dic1920161.pdf)
- SSPD. Superintendencia de Servicios Públicos Domiciliarios. (2019). Informe Sectorial de la Actividad de Aprovechamiento 2018. SSPD. -disponible en: [https://www.superservicios.gov.co/sites/default/archivos/Publicaciones/Publicaciones/2020/Ene/informe\\_sectorial\\_aprovechamiento\\_2018.pdf](https://www.superservicios.gov.co/sites/default/archivos/Publicaciones/Publicaciones/2020/Ene/informe_sectorial_aprovechamiento_2018.pdf)
- Suyadi,-Manullang, C.Y. (2020). Distribution of plastic debris pollution and its implications on mangrove vegetation. *Marine Pollution Bulletin*, 160(2020), art. 111642. DOI: <https://doi.org/10.1016/j.marpolbul.2020.111642>
- UNEP. United Nations Environment Programme. (2009). Sustainable coastal tourism: an integrated planning management approach. (Issue August 2016). United Nations Environment Programme, Milan. Available at: <https://wedocs.unep.org/handle/20.500.11822/7819>
- Valiela, I., Bowen, J.L., and York, J.K. (2001). Mangrove forests: one of the world's threatened major tropical environments. *BioScience*, 51, pp. 807–815. DOI: [https://doi.org/10.1641/0006-3568\(2001\)051\[0807:MFOOTW\]2.0.CO;2](https://doi.org/10.1641/0006-3568(2001)051[0807:MFOOTW]2.0.CO;2)
- Vélez-Mendoza, A. (2022). Marine litter in mangroves: composition, magnitude, and impacts. *Boletín de Ciencias de la Tierra*, (51), pp. 50–60. DOI: <https://doi.org/10.15446/rbct.101510>
- Vera, L.A. y De La Rosa-Muñoz, J. (2003). Estructura de la comunidad íctica de la ciénaga de mallorquín, Caribe Colombiano. *Boletín de Investigaciones Marinas y Costeras*, 32, pp. 231–242. DOI: <https://doi.org/10.25268/bimc.invemar.2003.32.0.268>
- Villate, D.A., Sánchez, H., Portz, L., Portantiolo, R., Bolívar-Anillo, H.J., and Anfuso, G. (2020). Mangrove forests evolution and threats in the Caribbean Sea of Colombia. *Water*, 12, art. 1113. DOI: <https://doi.org/10.3390/w12041113>
- Vivas-Aguas, L.J., Ibarra, K., Sánchez, J., Martínez, M., Nieto, Y., Moreno, Y., Cuadrado, I., Obando, P., Garces, O., Sánchez, D., Villarraga, M. y Sierra, O. (2015). Diagnóstico y evaluación de la calidad de aguas marinas y costeras en el Caribe y Pacífico colombianos. Informe Técnico 2014. Serie de Publicaciones Periódicas del INVEMAR, 4, 320 P. Disponible en: [http://www.invemar.org.co/documents/10182/14479/Informe+REDCAM+2015\\_mayo2016.pdf](http://www.invemar.org.co/documents/10182/14479/Informe+REDCAM+2015_mayo2016.pdf)
- Williams, A.T., Rangel-Buitrago, N., Anfuso, G., Cervantes, O., and Botero, C. (2016a). Waste impacts on scenery and tourism on the Colombian north Caribbean coast. *Tourism Management*, 55, art. 209e224. DOI: <https://doi.org/10.1016/j.tourman.2016.02.008>
- Williams, A.T., Randerson, P., Di Giacomo, C., Anfuso, G., Macias, A., and Perales, J. A. (2016b). Distribution of beach waste along the coastline of Cádiz, Spain. *Marine Pollution Bulletin*, 107(1), pp. 77–87. DOI: <https://doi.org/10.1016/j.marpolbul.2016.04.015>
- Williams, A.T., and Rangel-Buitrago, N. (2019). Marine waste: Solutions for a major environmental problem. *Journal of Coastal Research*, 35(3), pp. 648–663. DOI: <https://doi.org/10.2112/JCOASTRES-D-18-00096.1>
- A.J.L. Vélez-Mendoza;** is Biologist from the Universidad del Atlántico and current student of the Master of Sciences - Biology, Marine Biology Line of the Universidad Nacional de Colombia, CECIMAR Caribbean headquarters. In the thematic area of contamination by marine litter, he has experience in the course of the United Nations Environment Program and Open University MOOC on marine litter, with outreach participation with the community (conversations and symposiums), scientific events (eg: ACIMAR, SENALMAR, and RedCOLSI) and publications of scientific articles on ecosystems such as beaches, dunes and estuarine ecosystems submitted on *Marine Pollution Bulletin*, *Ocean and Coastal Management*, and *Estuarine Coastal and Shelf Science*.  
ORCID: 0000-0003-3878-8107
- C. Villamil;** is Marine Biologist with MSc., diplomas in Coastal Marine Integrated Management; and Ecological Restoration. More than 17 years of experience in formulation, execution, and coordination of planning, management, and environmental projects, prioritizing multi-stakeholder relationships. Advisor and consultant for various entities, especially environmental authorities. Leader of multiple processes of passive restoration and monitoring of mangroves in Colombia. Author of the Guide to Ecological

Restoration of Mangroves in Colombia (GREM). Participation in projects with national and foreign entities, to understand the environmental, economic, and social effects of extreme weather events El Niño/La Niña in Colombian mangrove ecosystems. Technical advisor in management, strategic planning, and multi-stakeholder relations, promoting dialogue and joint construction from the territories.

ORCID: 0000-0003-0944-6001

**K.I. Castellanos-Romero;** is Biologist, MSc in Environmental Sciences, with a PhD. in Biological Sciences. I'm an associate professor, attached to the Faculty of Basic Sciences, for 17 years and with 19 years of experience in research. Her research trajectory focuses on the evaluation of water and sediment quality in lentic systems of the Colombian Caribbean, where she was involved in biological quality with the use of bioindicators; acting as principal investigator. Currently, she is co-leader of the Colombian Caribbean Biodiversity Research Group (Recognized A1 MinSciences) and researcher of the Interdisciplinary Group of Marine and Environmental Sciences GICMARA (C MinSciences). These works involved applications in areas such as Ecology and Biotechnology; I'm the coordinator of environmental projects for public and private companies, environmental authorities, and territorial entities.

ORCID: 0000-0001-5215-3285

**Y. del C. Domínguez-Haydar;** is Biologist, with a MSc. in Biology, a MSc. in Ecosystem Restoration, and a PhD. in Ecology, Conservation, and Ecosystem Restoration. With 17 years of professional and research experience and 15 years of teaching experience. Member of the Colombian Network of Ecological Restoration. Competencies include biological data management and experimental design, as well as the interpretation of processes related to ecological restoration, mainly using soil macrofauna, with an emphasis on ants as bioindicators. Other interests include soil ecology and mangrove ecology.

ORCID: 0000-0002-0774-8391